Renewable Energy 126 (2018) 14-20

Contents lists available at ScienceDirect

Renewable Energy

journal homepage: www.elsevier.com/locate/renene

Two-stage anaerobic digestion of the organic fraction of municipal solid waste – Effects of process conditions during batch tests

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A R T I C L E I N F O

Article history: Received 30 May 2017 Received in revised form 22 January 2018 Accepted 15 March 2018 Available online 16 March 2018

Keywords: Organic waste Dark fermentation Anaerobic digestion Hydrogen Methane Biogas

ABSTRACT

Two-stage anaerobic digestion (AD) batch tests were performed using the organic fraction of municipal solid waste as substrate. Effects of different combination of initial pH (5.5, 7, and 9) and food to microorganism (F/M) ratio (from 0.5 to 6 gVS/gVS) were investigated for hydrogen and methane productions during the first and the second stage of AD, respectively.

Results showed that both initial pH and F/M ratio had an impact on hydrogen yield, hydrogen production rate and duration of lag phase. The highest hydrogen yield of 29.8 mLH₂/gVS was obtained at initial pH of 5.5 and F/M ratio of 6. However, the highest hydrogen production rate (65 mLH₂/gVS/d) was recorded at pH of 9 and F/M ratio of 6. Increasing the initial pH from 5.5 to 9, led to shorter lag phases for all F/M ratios. Methane production from second phase was not significantly influenced by the F/M ratios tested in the first digestion phase. When compared to single-phase AD, two-stage AD tests resulted in enhanced methane production rates from 37.3 to 68.5 mLCH₄/gVS/d, reducing by half both the lag phase and the time required to reach maximum methane production.

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1. Introduction

Two-stage anaerobic digestion process has recently been suggested as an option to maximize the amount of energy recoverable from biodegradable organic waste in terms of hydrogen (H₂) and methane (CH₄) [1–3]. H₂ is a clean energy carrier and has a highenergy density. H₂ has, in fact, the highest calorific value among other fuels and its combustion does not lead to carbon emissions. H₂ can be produced from cheap organic wastes and wastewaters in a process called dark fermentation [3,4]. Biomethane can play a central role in the development of the circular economy principle. It is a source of energy that can be used for power and heat production but also as a gaseous vehicle fuel, it can replace natural gas and be fed into national gas grids or be used as a feedstock for producing chemicals and materials [5].

Pre-treatments are often applied to enhance biogas productivity

ability. During the fermentation stage of AD, organic substances are hydrolysed and converted to H₂ and volatile fatty acids (VFAs) by hydrogen producing bacteria. Optimisation of the H₂ production phase can lead to an improved hydrolysis and therefore higher energetic exploitation of waste materials. The advantages traditionally indicated for two-stage digestion systems, if compared to single stage AD, are shorter substrate retention time, enhanced solids degradation efficiencies [8–10], enhanced hydrolysis with a subsequently higher CH₄ production

of substrates [6,7] and fermentation step for H₂ production itself could be seen as a pre-treatment to increase overall biodegrad-

[11–13] and potentially higher organic loading rates [14]. Despite these advantages, the higher complexity of two stage digestion plants, if compared to single digestion, limited the diffusion of this option to less than 10% of current digestion capacity [15].

The possibility of simultaneous H_2 and CH_4 productions from the same feedstock, rather renovates the interest of this kind of plant configuration and this option is currently receiving growing interest with several investigations at lab and pilot scale level [16–18]. Besides reaching higher energy yields, two-stage AD promotes a stronger bio-stabilisation of the treated organic waste [19,20] and could also lead to the production of metabolites to be





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used as renewable and biodegradable substitutes for petrochemical products [21,22].

The main variables influencing both single and two-stage AD performances are substrate C/N ratio, reactors retention time, inoculum, pH, and food to microorganism ratio (F/M). Optimal substrate C/N ratio for single stage AD was found to be between 15 and 30 while substrates with C/N ratios lower than 10 should be treated only in a two-stage AD process [23]. Substrate retention time is generally short during the H₂ production phase (20 h-4 days) to avoid the risk of methanogenic activity even though excessively short retention times may be detrimental for substrates characterised by slow hydrolysis rates [24,25]. In contrast, a longer retention time is needed for CH₄ production (20-30 days) in order to reach complete substrate degradation and enhanced digestate stabilization [26,27]. Pure or mixed microflora cultures can be used for H_2 production from single or two-stage digestion. Generally, H_2 yields are higher when pure cultures are specifically chosen accordingly to the fermented substrate while mixed cultures could show better adaptability towards environmental stress, nutrients availability and process conditions [28,29].

Notwithstanding, there is still the need to define optimal operational parameters and procedures to promote the successful succession of the two phases without compromising operational condition for the methanogenic stage. In particular, there is a considerable lack of comprehensive studies relating to the effects produced by initial operational parameters of fermentation on the second phase of the process.

Data on H_2 and CH_4 production yields reported in scientific studies on two-stage AD process are illustrated in Table 1. Results indicate that generally there is a good energy recovery potential from the treatment of organic waste suggesting that efforts in assessing and proving the advantages of different operational conditions as well as of the energy recovery potential can stimulate the application of two-stage AD and diffuse the production of renewable H_2 and CH_4 from organic residues.

The aim of this study was to investigate the effect of F/M ratio and initial pH on hydrogen and methane productions in a twostage anaerobic digestion process using organic fraction of municipal solid waste (OFMSW) as substrate.

2. Materials and methods

2.1. Substrate and inoculum

OFMSW samples were collected from the waste receiving area of an anaerobic digestion plant treating organic waste located in Padova, Italy. The OFMSW delivered at the plant is source segregated at household level and the collection area involves a population of about 130,000 inhabitants. Samples were properly sorted and stored before use [30]. Samples were chopped with a food grinder and diluted with water at a ratio 1:2 (kg/L) prior to use as substrate lab scale tests. Granular sludge collected from a full-scale Upflow Anaerobic Sludge Blanket (UASB) digester of a brewery located in Padova, was used as inoculum (mixed culture). OFMWS and sludge samples were characterized for the following parameters: Total Solids (TS), Volatile Solids (VS), Total Carbon (TC), Total Kjeldahl Nitrogen (TKN), and Chemical Oxygen Demand (COD) (Table 2).

2.2. Two-stage digestion batch tests

Experimental design was planned in order to study the combination effects of each investigated initial pH and F/M ratio on twostage AD process (Table 3). The following conditions F/M and pH were tested during the first stage of digestion for H₂ production. F/ M ratios were 0.5, 1, 2, 4, and 6 gVS/gVS. Initial pH values were 5.5, 7, and 9.

Two-stage digestion batch tests were carried out using 1L glass bottles sealed with a silicon plug and a working volume of 500 mL. Different F/M ratios were achieved changing the amount of inoculum in each test while substrate concentration was kept constant at 5 gVS/L. MES ($C_6H_{13}NO_4S$) was used to obtain an initial pH of 5.5, while (sodium carbonate, Na_2CO_3) was used to reach an initial pH value of 9. No buffer was used for the initial tests at neutral pH (7.0).

For first AD stage (fermentation tests), the inoculum was thermally pre-treated for 4 h at 100 °C to inhibit methanogenic archaea and to enhance the activity of hydrogen producing bacteria in the mixed culture [31]. For the second AD stage, the same amount of sludge was added in each bottle in order to obtain the same F/M ratio which was determined by dividing the original F by the new M. To promote CH₄ production during the second AD stage (methane production), F/M ratio was fixed at 0.5 gVS/gVS, while initial pH was set at 8.5 by dosing Na₂CO₃.

A single-stage AD test (methane production only), characterized by a F/M ratio equal to 0.5 gVS/gVS and a pH of 8.5, was run in parallel in order to compare methane production yields with hydrogen and methane yields obtained through the two-step process.

The bottles were flushed with N₂ gas for 3 min to ensure anaerobic conditions and incubated at a temperature of 35 ± 1 °C. Incubation lasted 45 days for two-stage AD tests and 60 days for single-stage ones. All tests were performed in duplicate.

2.3. Analytical methods

TS, VS, COD, TKN and alkalinity were analysed according to standard methods [32]. TC was analysed by a TOC analyser (TOC-V CSN, Shimadzu). The volume of biogas produced during two-phase digestion tests was measured by means of the water displacement method [33]. H₂, CO₂, and CH₄ concentrations in biogas were measured by a gas chromatograph (HP5890) equipped with thermal conductivity detector (TCD), HP-MOLSIV and HP-PLOT U columns, and nitrogen as carrier gas.

 H_2 and CH_4 volumes produced in the time interval between each measurement [t - (t-1)] were calculated using a model taking into consideration the gas concentration at time t and time t-1, together with the total volume of biogas produced at time t, the concentration of specific gas at times t and t-1, and the volume of head space of reactors [34]. The following equation was applied:

$$V_{C,t} = C_{C,t} * V_{G,t} + V_{H} * (C_{C,t} - C_{C,t-1})$$
(1)

where $V_{C,t}$ – hydrogen or methane volume produced in the interval between t and t - 1; $C_{C,t}$, $C_{C,t-1}$ – hydrogen or methane concentrations measured at times t and t-1; $V_{G,t}$ – volume of gas produced between time t and t-1; V_H – volume of the headspace of reactors.

To compare the results obtained from the batch tests, data were interpolated on the basis of the Gompertz model [35]. The Gompertz mathematical expression is described in Equation (2):

$$P(t) = P_{\max}e\left\{-e\left[\frac{R^*e}{P_{\max}}\right](\lambda - t) + 1\right\}$$
(2)

where P (t) is the cumulated H₂ or CH₄ production at time t; P_{max} is the maximum H₂ or CH₄ production; R is the maximum production rate; and λ is the lag phase. The results related to production rate (R) and duration of the lag phase (λ) were applied to compare the different investigated operative conditions.

Data on H_2 and CH_4 productions are expressed at a temperature of 0 °C and pressure of 1 atm (Normal conditions).

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Substrate	Hydrogen potential production (mLH ₂ /gVS)	Methane potential production (mLCH ₄ /gVS)	Reference
Dairy processing waste	40.15	34.2	[52]
Kitchen waste	36	135	[53]
OFMSW	43	500	[13]
OFMSW	90	560	[54]
Potato residues	31	387	[55]
Steam-peeling potato waste	134 ^a	183 ^a	[56]
Common wheat waste	47 ^a	202 ^a	[56]
Vinegar residue	53.2	192	[57]

^a mL/gCOD.

Table 2

Average substrate and inoculum characteristics.

Parameter	OFMSW	Granular sludge
TS (%)	75	15
VS (%TS)	90	53
TC (%TS)	50.2	29.6
TKN (gN/kgTS)	8.7	43
C/N (gC/gN)	58	7
COD (gO ₂ /kgTS)	300	693

Table 3

Initial operational conditions of two-stage and single stage batch tests.

Run	First stage, ferme	ntation	Second stage, me	thanization
	F/M (gVS/gVS)	Initial pH	F/M (gVS/gVS)	Initial pH
А	0.5	5.5	0.5	8.5
В	1.0	5.5	0.5	8.5
С	2.0	5.5	0.5	8.5
D	4.0	5.5	0.5	8.5
Е	6.0	5.5	0.5	8.5
F	0.5	7.0	0.5	8.5
G	1.0	7.0	0.5	8.5
Н	2.0	7.0	0.5	8.5
Ι	4.0	7.0	0.5	8.5
L	6.0	7.0	0.5	8.5
М	0.5	9.0	0.5	8.5
Ν	1.0	9.0	0.5	8.5
0	2.0	9.0	0.5	8.5
Р	4.0	9.0	0.5	8.5
Q	6.0	9.0	0.5	8.5
Cimala				

Single stage, methane production

	F/M (gVS/gVS)	Initial pH
R	0.5	8.5

3. Results and discussion

3.1. Effect of F/M ratio and initial pH during the first AD stage – fermentation

Hydrogen production yields obtained during the first stage (fermentation) are shown in Table 4. Hydrogen yields were slightly lower than the values reported in the literature for similar substrates (Table 1). This could be due to either a decreased hydrolytic activity after the long anaerobic sludge pre-treatment or the specific substrate composition used in this study. A decreased hydrolytic activity could be a side effect of the heating process at 100 °C for 4 h. Shah et al. [36] assessed the viability of isolates from granular sludge after a pre-treatment similar to the one applied in this study (2 h and 4 h at 100 °C) and observed that isolates still active after the heat shock exhibited a broad range of hydrolytic activities. It is, therefore, presumable that the slightly lower H₂ yields compared to those from similar studies could be due to the specific OFMSW composition. It was observed that yields of H₂

production from OFMSW collected in different seasons varied during the year due to changes in the OFMSW composition [30]. Various studies also confirmed that carbohydrate content of organic wastes directly affects H₂ production suggesting that a lack of fractions rich in sugars or starch could reduce the H₂ productions via biological fermentation [3,37,38].

In general, two days were enough to complete hydrogen production but a total of four days was waited to ascertain the plateau and no methane production was observed during the fermentation tests, indicating that the sludge pre-treatment was effective in inhibiting methanogens. Moreover, none of the conditions tested during first phase, in terms of F/M ratios and pH, favoured the reactivation of methanogens even after the fermentation stopped (data not shown).

Results of the data modelling with Gompertz equation (2) are reported in Table 4 and plots of H₂ yields vs. F/M ratios and of lag phase duration (λ) vs. initial pH are shown in Fig. 1a and Fig. 1b, respectively. The additional parameter, *t*₉₅, defined as the time required for H₂ production to attain 95% of the total cumulative yield [24], was also calculated and reported in Table 4. The highest H₂ yield of 29.8 mLH₂/gVS was recorded from test E, characterised by a F/M ratio of 6 and an initial pH of 5.5. This test was also characterised by a lag phase of 14.9 h and a maximum production rate of 40.3 mLH₂/gVS/d. The lowest H₂ yield was recorded from test F, characterised by a F/M ratio of 0.5 and an initial pH of 7.0 (Table 1). The H₂ production rate for this test was also the lowest (12.9 mLH₂/gVS/d). Whilst the low yield and production rate, the lag phase for this test was shorter than that for test E.

In general, high H₂ yields were observed from tests with F/M ratios of 4 and 6 gVS/gVS (Fig. 1a) and short lag phases were observed for tests with an initial pH of 7 or 9 (Fig. 1b). Production rates (R) were influenced neither by F/M ratio nor by initial pH, even though faster production rates (R) are generally associated with higher production yields (P_{max}).

In the present study, different F/M ratios were obtained by changing the sludge concentration in the reactors while substrate concentration was kept constant. Despite the presumable larger presence of hydrogen producing bacteria at lower F/M ratios, this condition did not lead to higher H₂ yields. A larger variability in bacterial populations present in the mixed microflora with low F/M ratios could have introduced also non-hydrogen forming bacteria competing for the same substrates or hydrogen consuming bacteria and this could have had a measurable impact on hydrogen yields. The higher F/M ratios, on the contrary, were obtained by lower biomass concentrations and this condition could have reduced the possibility of non-hydrogen forming bacteria or hydrogen consuming bacteria to have an effect on hydrogen yields. Alibardi et al. [31] indicated that long heat pre-treatments strongly influence microbial viability, with reductions of order of magnitudes of active bacteria levels. Despite this effect, high bacterial concentrations could allow niches of non-hydrogen forming or hydrogen consuming bacteria to grow sufficiently to produce an effect on

Run	First stage, fermentation	First stage, fermentation			Modelling results			
	Hydrogen production (mLH ₂ /gVS)	Final pH	R (mLH ₂ /gVS/d)	λ (h)	P _{max} (mLH ₂ /g VS)	<i>t</i> ₉₅ (d)		
A	18.0	6.0	55.5	14.2	18.0	1.1		
В	13.0	6.0	20.6	11.2	13.0	1.4		
С	14.2	6.0	25.3	11.5	14.2	1.3		
D	16.4	6.0	19.4	10.2	16.4	1.7		
Е	29.8	5.5	40.3	14.9	29.8	1.7		
F	7.0	6.0	12.9	9.0	7.0	1.2		
G	15.9	5.0	50.0	9.6	15.9	0.9		
н	10.2	5.0	18.0	5.3	10.2	1.0		
I	28.2	5.0	54.2	7.2	28.8	1.1		
L	17.5	5.0	64.8	7.2	17.5	0.7		
М	7.9	6.5	27.3	8.5	7.9	0.8		
N	14.9	5.5	55.3	8.1	14.9	0.7		
0	24.0	5.0	60.0	4.8	24.0	0.8		
Р	24.3	5.0	64.8	6.0	24.3	0.8		
Q	23.6	5.0	65.0	5.9	23.6	0.8		

 Table 4

 Hydrogen production yields (average values), final pH at the end of the first stage of AD batch tests and results of the data modelling with Gompertz equation (2).

hydrogen yield. On the contrary, when biomass concentrations are kept low, these niches are not able to influence overall hydrogen productions that are only defined by the activity of fast growing hydrogen producing bacteria. F/M ratio has therefore a direct effect on microbial activities of different populations present in the mixed microflora and to maximise H₂ production, small concentrations are sufficient to obtain efficient hydrogen conversions [31]. Pan et al. [39] investigated how F/M ratio affects H₂ production from food waste under mesophilic and thermophilic conditions but with no pre-treatment to enhance hydrogen production. Differently from the approach in the present research study, a constant biomass concentration was used by Pan et al. [39] while substrate concentration was changed. Optimal F/M ratios of 6 and 7 were identified under mesophilic and thermophilic conditions, respectively. Low F/M ratios (<3) led to high methane productions and at high F/M ratios (>7) low H₂ yields were observed. These results confirm how an optimal balance between biomass and substrate concentrations needs to be identified to enhance hydrogen production and avoid the activity of other bacterial species not contributing or negatively affective hydrogen fermentation.

Initial pH and F/M ratio also influenced hydrogen production rate (R) and lag phase duration (λ) (Table 4). When initial pH was increased from 5.5 to 9, a shorter lag phase was observed for all F/M ratios (Fig. 1b). The longest lag phase (14.9 h - Test E) was observed with F/M ratio and initial pH of 6 and 5.5, respectively. The shortest lag phase (4.8 h - Test O) corresponded to F/M ratio and initial pH of 2 and 7, respectively. These results suggest that a neutral to basic pH could speed up the activity of hydrogen forming bacteria after the heat pre-treatment while an initial acid condition imposes a longer acclimating phase before hydrogen production starts. These results



Fig. 1. Distribution of maximum hydrogen (H_2) production (P_{max}) over F/M ratio (a) and lag phase (λ) over initial pH (b).

are in accordance with Chen et al. [35] who reported longer lag phases when mixed microflora inoculum was cultivated at pH of 5 (compared to pH 6 and 7) after an enrichment phase at both acid or basic conditions. Similarly, Ferchichi et al. [40] reported a significantly long lag phase of 43.26 h with an initial pH of 5 and a short lag phase of 3.06 h when the pH was 8, using cheese whey as substrate. The initial low pH conditions can result in the protonation of weak acids contained in cheese whey, which may pass freely through the cell's membrane into its cytoplasm causing its consequent acidification [41]. This internal condition could result in loss of activity by the glycolytic enzymes and structural damage of the cell membrane that can lead to prolonged re-activation phases after external stresses to the inoculum and, consequently, longer lag phases [40]. The low pH values set by using MES in this study, could have led to a similar effect and produced the observed delay (Table 4).

Operational pH is also one of the key factors in dark fermentative H₂ production. It may affect hydrogenase activity and metabolic pathways towards different by-products generation [42]. In all tests, pH decreased to values between 5.5 and 6 at the end of the fermentation (Table 4). These results indicate that, despite the different pH set at the beginning of the tests, the fermentation products established an acid environment even at high initial pH conditions (pH = 9). Optimal initial pH of 5.5–6.0 has been reported by many studies for mesophilic dark fermentation [3,40,43–45]. Low pH (4.5–6) leads to a higher concentration of acetic and butyric acids which are soluble metabolites whose production pathways are accompanied by H₂ production [46]. Moreover, the activities of H₂ consuming microorganisms like methanogens, homoacetogens, and propionic acid bacteria decrease at low pH conditions [42,47]. This study also demonstrated that high initial pH speeded up the inoculum reactivation with short lag phases. It is, therefore, presumable that an optimal combination of initial pH and operational pH during the fermentation process, could enhance the overall hydrogen production by combining short lag phases with high hydrogen yields. Further studies are anyway required to confirm or rebut this hypothesis.

3.2. Effect of F/M ratio and initial pH during the second AD stage – methane production

Methane production yields during the second AD stage are reported in Table 5. The highest methane production of 620 mLCH₄/ gVS was obtained from test F, while the lowest was measured from

test D (463 mLCH₄/gVS). The average methane production of 544 mLCH₄/gVS was obtained from all the tests at various F/M ratios and initial pH conditions. The maximum methane production from test R, carried out in a single digestion phase for methane production, resulted 633 mLCH₄/gVS. The lower methane yields obtained from the double digestion process could be explained by the fact that hydrogen was produced in the first digestion phase. A portion of the total electrons released by the biodegradation process was already passed to H₂ and therefore a reduction of the total methane production from the second phase could be expected. Notwithstanding, the additional amounts of methane producible according to stoichiometry (4H₂ + CO₂ \rightarrow CH₄ + 2H₂O) are only 7.45 and 1.75 mLCH₄/gVS from the maximum and minimum H₂ production yields, respectively.

The outputs from the second digestion phase (Tests A to Q) and the single digestion process (Test R) displayed a similar pattern although the lag phase for single phase digestion was longer and time required to reach maximum methane production was almost doubled (Table 6). Indeed, for two-stage AD, hydrolysis and acidogenesis occur during the first stage resulting in enhanced VFAs production which can be converted to CH₄ rapidly during the second stage [12,13,48]. The optimal conditions for hydrolytic bacteria and methanogenic archaea may be different and splitting the process into two phases, provides the opportunity for the specific optimization of each phase. Differently, single-stage AD, for which hydrolysis is the rate limiting process, combines hydrolysis, acidogenesis and methanogenesis with a consequently longer lag phase than that of the second stage of a two-stage AD.

Average maximum methane productions (Fig. 2) decreased in line with an increase of the F/M ratio (applied in the first stage) up to 4. However, an opposite trend was observed when passing from a F/M ratio of 4–6, displaying a pattern similar to that observed for H₂ production during the first AD stage (Fig. 1).

Comparing trends in Figs. 1a and 2, it can be highlighted that lower hydrogen productions are associated with higher methane yields for all F/M ratios. In particular, test F, characterised by a F/M ratio of 0.5 and pH of 7, produced the lowest amount of hydrogen (7 mLH₂/gVS) and the highest amount of methane (619 mLCH₄/gVS). In accordance with Schievano et al. [49], single-stage Biochemical Methane Potential (BMP) outputs featured higher methane yields than those achieved from a two-stage AD process, although with a

Table 6

Comparison of Gompertz equation modelling results from single-stage and two-stage AD processes (average values between all tests). R - methane production rate, λ - lag phase, and t_{max} - time needed for maximum methane production.

	P_{max} (mLCH ₄ /gVS)	R (mLCH ₄ /gVS/d)	λ(d)	$t_{max}(d)$
Single-stage AD	633	37.3	12.0	40
Two-stage AD	544	68.5	5.4	20

longer lag phase and lower maximum production rate. The slightly lower CH₄ yields obtained for two-stage AD could be due to the previous recovery of H₂ which is also a substrate for methane production. In fact, in a single-stage AD, CH₄ could be obtained both from VFAs conversion by acetoclastic methanogens and by H₂ and CO₂ conversion by hydrogenotrophic methanogenic archaea. In contrast to our results, Voelklein et al. [48] reported a 23% increase in methane production from a two-stage AD of restaurant food waste rather than a single-stage process. Likewise, Liu et al. [13] recovered 21% more methane from two-stage AD tests performed on mixed organic waste.

The final pH of two-stage AD process ranged between 7.5 and 8, while for single stage AD process final pH resulted 7.0 (Table 5). For both the methane production phase of the two-stage AD and for single AD, initial pH was set at a value of 8.5. The slightly lower final pH observed for single AD could suggest that a higher buffer capacity is required for systems where all phases of digestion are carried out in one single reactor. On the contrary, for two-stage AD a lower buffer capacity is required as acidic fermentation residues from first stage are rapidly converted to CH₄. Notwithstanding, the results suggest that process condition and fermentation activity during first stage have an impact on second stage performance. Tests performed at F/M ratios of 4 and 6 (D, E, I, L, P, and Q) were in fact always related to the lowest final pH value, and therefore a higher buffer capacity, probably because of the higher biological metabolites production favoured by high F/M ratios [50].

The first AD stage, aimed at hydrogen production, may also be viewed as an effective pre-treatment for the subsequent production of methane, providing a VFA-rich substrate ready to be digested by methanogenic archaea. The average CH₄ yield from two-stage AD (544 mLCH₄/gVS) was lower than the one from single-stage AD (633 mLCH₄/gVS) (Tables 5 and 6). However, if similar conditions

Table 5

Methane production yields, final pH at the end of the second stage of AD batch tests and results of the data modelling with Gompertz equation (2).

Run	Second stage, methane production		Modelling results			
	Methane production (mLCH ₄ /gVS)	Final pH	R (mLCH ₄ /gVS/d)	λ (d)	P _{max} (mLCH ₄ /g VS)	<i>t</i> ₉₅ (d)
Α	527	8	101.4	7.9	517	15.3
В	550	8	76.4	5.4	549	15.9
С	499	7.5	68.0	5.4	493	16.0
D	489	7.5	52.8	5.0	463	17.8
Е	582	7.5	62.5	5.1	562	18.2
F	619	8	67.4	4.9	620	18.3
G	590	8	62.1	4.8	600	18.9
Н	532	8	56.6	4.7	541	18.7
I	474	7.5	50.1	4.2	482	18.3
L	523	7.5	53.1	4.2	534	18.9
М	606	8	94.9	6.8	614	16.2
Ν	566	8	73.2	5.9	573	17.3
0	554	8	75.2	5.8	553	16.5
Р	532	7.5	67.8	5.9	529	17.3
Q	529	7.5	66.6	5.6	529	17.2
	Single stage, methane production		Modelling results			
	Methane production (mLCH ₄ /gVS)	Final pH	R (mLCH ₄ /gVS/d)	λ (d)	P _{max} (mLCH ₄ /g VS)	t ₉₅ (d)
R	626.1	7.0	37.3	12.0	633	36.8



Fig. 2. Distribution of maximum methane (CH₄) production (P_{max}) obtained from the second phase over F/M ratios tested during the first phase.

were considered (F/M = 0.5), approximately equal yields were obtained from single-stage and two-stage AD (626.1 and 619 mLCH₄/ gVS, respectively). Methane production rate (R) doubled (Table 6), whilst both lag phase and time required to reach the maximum methane production were reduced by half when passing from single-stage to two-stage AD process. These findings are in accordance with Leite et al. [51] who achieved a 15% increase of produced energy from single-stage to two-stage AD system. In fact, when splitting the AD process into two-stages, the first stage may be regarded as a pre-treatment to increase the methane production rate and to shorten the lag phase, as confirmed by the results reported in the present paper. The faster production rate, accompanied by a shorter lag phase, proves a significant overall benefit of two-stage over single-stage AD. It is important to highlight that the maximum methane productions during the two-stage processes were reached after 20 days of incubation; on the contrary, for the single-stage test, the maximum methane production (626.1 mLCH₄/gVS) was reached after about 40 days. By comparing the potential energy output of the two processes, it is possible to highlight how a double phase digestion process could be energetically more favourable if compared to a single-phase digestion when the time for digestion (i.e. digester volume or solid retention time) is fixed at 22 days. In the single-stage AD test the cumulative methane production registered after 22 days of incubation was 366.2 mLCH₄/gVS. Considering a period of 2 and 20 days for hydrogen and methane productions, respectively, for a two-stage digestion, and of 22 days in the case of the single-stage process, the potential energy output for produced fuel gases is reported in Fig. 3. These choices were made on the basis of the average time required to reach maximum hydrogen and methane productions in the two-stage process. According to Fig. 3, all two-stage tests were energetically more favourable than single-stage tests. These results confirm that the implementation of a two-stage digestion processes for sequential H₂ and CH₄ production from OFMSW could enhance methanogenic phase performances and increase the overall potential energy production thank to faster digestion processes.

4. Conclusion

The present study investigated the effects of two parameters, initial pH and food to microorganism ratio, on hydrogen and methane productions obtained from the organic fraction of municipal solid waste in a two-stage AD process. Data analysis revealed how a variation in initial pH value influenced substrate degradation kinetics and total hydrogen production. Kinetics were



Fig. 3. Potential energy output from single-stage and two-stage tests after 22 days of digestion (2 days first stage, 20 days second stage). H₂ energy density = $120 \text{ MJ/kg} - CH_4$ energy density = 50 MJ/kg (Single-stage AD yielded 366.2 mLCH₄/gVS after 22 days of digestion).

favoured by initial alkaline conditions (pH = 9) linked to faster production rates and shorter lag phase. High F/M ratios were found to facilitate hydrogen production, with the most favourable condition being identified at a F/M ratio of 6. Peak methane production (619 mLCH₄/gVS) recorded during the second AD stage of BMP test characterized by a F/M ratio of 0.5 and an initial pH of 7, was close to the value of 633 mLCH₄/gVS obtained during the single-stage process. There was no evident relationship between initial pH values during fermentation and methane production, probably due to pH adjustment performed on completion of fermentation tests, while an increase in F/M ratio from 0.5 to 4 resulted in a slight decrease in methane production. The fermentation phase, in addition to promoting hydrogen recovery, represents an efficient means of pre-treatment aimed at enhancing subsequent methane production. In comparison with the single-stage AD process, a twostage process elicits faster methane production, a shorter lag phase, and a better energetic exploitation of OFMWS, as demonstrated by the achieved energy output.

References

- B. Ruggeri, T. Tommasi, G. Sassi, Energy balance of dark anaerobic fermentation as a tool for sustainability analysis, Int. J. Hydrogen Energy 35 (19) (2010) 10202–10211.
- [2] F. Girotto, W. Peng, R. Rafieenia, R. Cossu, Effect of aeration applied during different phases of anaerobic digestion, Waste Biomass Valorization (2016) 1–14, https://doi.org/10.1007/s12649-016-9785-9.
- [3] R. Rafieenia, M.C. Lavagnolo, A. Pivato, Pre-treatment technologies for dark fermentative hydrogen production: current advances and future directions, Waste Manag. (2017), https://doi.org/10.1016/j.wasman.2017.05.024.
- [4] Z. Ma, C. Li, H. Su, Dark bio-hydrogen fermentation by an immobilized mixed culture of Bacillus cereus and Brevumdimonas naejangsanensis, Renew. Energy 105 (2017) 458–464.
- [5] P. Weiland, Biogas production: current state and perspectives, Appl. Microbiol. Biotechnol. 85 (4) (2010) 849–860.
- [6] Q. Zhang, J. Hu, D. Lee, Biogas from anaerobic digestion processes: research updates, Renew. Energy 98 (2016) 108–119.
- [7] R. Rafieenia, F. Girotto, W. Peng, R. Cossu, A. Pivato, R. Raga, M.C. Lavagnolo, Effect of aerobic pre-treatment on hydrogen and methane production in a two-stage anaerobic digestion process using food waste with different compositions, Waste Manag. (2016), https://doi.org/10.1016/ j.wasman.2016.10.028.
- [8] H. Nguyen, S. Heaven, C. Banks, Energy potential from the anaerobic digestion of food waste in municipal solid waste stream of urban areas in Vietnam, Int. J. Energy 5 (4) (2014) 365–374.
- [9] E.T. Brummeler, M.M.J. Aarnink, I.W. Koster, Dry anaerobic digestion of solid organic waste in a biocel reactor at pilot-plant scale, Water Sci. Technol. 25 (1992) 301–310.
- [10] S. Ghosh, J.P. Ombregt, P. Pipyn, Methane production from industrial wastes by two-phase anaerobic digestion, Water Res. 19 (1985) 1083–1088, https://

doi.org/10.1016/0043-1354(85)90343-4.

- [11] J. Massanet-Nicolau, R. Dinsdale, A. Guwy, Utilising biohydrogen to increase methane production, energy yields and process efficiency via two stage anaerobic digestion of grass, Bioresource 189 (2015) 379–383.
- [12] O. Pakarinen, P. Kaparaju, J. Rintala, Hydrogen and methane yields of untreated, water-extracted and acid (HCl) treated maize in one-and two-stage batch assays, Int. J. 36 (22) (2011) 14401–14407.
- [13] D. Liu, D. Liu, R.J. Zeng, I. Angelidaki, Hydrogen and methane production from household solid waste in the two-stage fermentation process, Water Res. 40 (2006) 2230–2236, https://doi.org/10.1016/j.watres.2006.03.029.
- [14] J. Ariunbaatar, E. Scotto Di Perta, A. Panico, L. Frunzo, G. Esposito, P.N.L. Lens, F. Pirozzi, Effect of ammoniacal nitrogen on one-stage and two-stage anaerobic digestion of food waste, Waste Manag. 38 (2015) 388–398, https:// doi.org/10.1016/j.wasman.2014.12.001.
- [15] L. De Baere, B. Mattheeuws, Anaerobic digestion of MSW in Europe, Biocycle 51 (2010) 24–26.
- [16] Z. Zuo, S. Wu, W. Zhang, R. Dong, Effects of organic loading rate and effluent recirculation on the performance of two-stage anaerobic digestion of vegetable waste, Bioresour. Technol. 146 (2013) 556–561.
- [17] D.-Y. Lee, Y. Ebie, K.-Q. Xu, Y.-Y. Li, Y. Inamori, Continuous H2 and CH4 production from high-solid food waste in the two-stage thermophilic fermentation process with the recirculation of digester sludge, Bioresour. Technol. 101 (2010) S42–S47, https://doi.org/10.1016/j.biortech.2009.03.037.
- [18] M. Park, J. Jo, D. Park, D. Lee, J. Park, Comprehensive study on a two-stage anaerobic digestion process for the sequential production of hydrogen and methane from cost-effective molasses, Int. J. Hydrog. 35 (12) (2010) 6194–6202.
- [19] J. Lim, C. Chen, I. Ho, J. Wang, Study of microbial community and biodegradation efficiency for single-and two-phase anaerobic co-digestion of brown water and food waste, Bioresour. Technol. 147 (2013) 193–201.
- [20] L. Wu, A. Higashimori, Y. Qin, T. Hojo, K. Kubota, Comparison of hyper-thermophilic-mesophilic two-stage with single-stage mesophilic anaerobic digestion of waste activated sludge: process performance and microbial community analysis, Chem. Eng. 290 (2016) 290–301.
- [21] F. Girotto, L. Alibardi, R. Cossu, Food waste generation and industrial uses: a review, Waste Manag. 45 (2015) 32–41, https://doi.org/10.1016/ j.wasman.2015.06.008.
- [22] F. Girotto, A. Pivato, R. Cossu, G.E. Nkeng, M.C. Lavagnolo, The broad spectrum of possibilities for spent coffee grounds valorisation, J. Mater. Cycles Waste Manag. (2017) 1–7, https://doi.org/10.1007/s10163-017-0621-5.
- [23] P. Weiland, One- and two-step anaerobic digestion of solid agroindustrial residues, Water Sci. Technol. 27 (1993).
- [24] G. De Gioannis, M. Friargiu, E. Massi, A. Muntoni, Biohydrogen production from dark fermentation of cheese whey: influence of pH, Int. J. 39 (36) (2014) 20930–20941.
- [25] A. Karlsson, L. Vallin, J. Rgen Ejlertsson, Effects of Temperature, Hydraulic Retention Time and Hydrogen Extraction Rate on Hydrogen Production from the Fermentation of Food Industry Residues and Manure, (n.d.). doi:10.1016/ j.ijhydene.2007.10.055.
- [26] D.E. Algapani, W. Qiao, F. Di Pumpo, D. Bianchi, S.M. Wandera, F. Adani, R. Dong, Long-term Bio-H 2 and Bio-CH 4 Production from Food Waste in a Continuous Two-stage System: Energy Efficiency and Conversion Pathways, 2017, https://doi.org/10.1016/j.biortech.2017.05.164.
- [27] L. Ding, J. Cheng, D. Qiao, L. Yue, Y.-Y. Li, J. Zhou, K. Cen, Investigating Hydrothermal Pretreatment of Food Waste for Two-stage Fermentative Hydrogen and Methane Co-production, 2017, https://doi.org/10.1016/ j.biortech.2017.05.114.
- [28] F.R. Hawkes, R. Dinsdale, D.L. Hawkes, I. Hussy, Sustainable fermentative hydrogen production: challenges for process optimisation, Int. J. Hydrogen Energy 27 (2002) 1339–1347. www.elsevier.com/locate/ijhydene (accessed September 27, 2017).
- [29] A. Shah, L. Favaro, L. Alibardi, L. Cagnin, A. Sandon, Bacillus sp. strains to produce bio-hydrogen from the organic fraction of municipal solid waste, Appl. Energy 176 (2016) 116–124.
- [30] L. Alibardi, R. Cossu, Composition variability of the organic fraction of municipal solid waste and effects on hydrogen and methane production potentials, Waste Manag. 36 (2015) 147–155, https://doi.org/10.1016/ j.wasman.2014.11.019.
- [31] L. Alibardi, L. Favaro, M.C. Lavagnolo, M. Basaglia, S. Casella, Effects of heat treatment on microbial communities of granular sludge for biological hydrogen production, Water Sci. Technol. 66 (2012), https://doi.org/10.2166/ wst.2012.336.
- [32] APHA/AWWA/WEF, Standard Methods for the Examination of Water and Wastewater, 2012, ISBN 9780875532356.
- [33] M.C. Lavagnolo, F. Girotto, O. Hirata, R. Cossu, Lab-scale co-digestion of kitchen waste and brown water for a preliminary performance evaluation of a decentralized waste and wastewater management, Waste Manag. (2017), https://doi.org/10.1016/j.wasman.2017.05.005.
- [34] S.W. Van Ginkel, S.E. Oh, B.E. Logan, Biohydrogen gas production from food processing and domestic wastewaters, Int. J. Hydrogen Energy 30 (2005) 1535–1542, https://doi.org/10.1016/j.ijhydene.2004.09.017.
- [35] C.C. Chen, C.Y. Lin, M.C. Lin, Acid-base enrichment enhances anaerobic

hydrogen production process, Appl. Microbiol. Biotechnol. 58 (2002) 224–228, https://doi.org/10.1007/s002530100814.

- [36] A.T. Shah, L. Favaro, L. Alibardi, L. Cagnin, A. Sandon, R. Cossu, S. Casella, M. Basaglia, Bacillus sp. strains to produce bio-hydrogen from the organic fraction of municipal solid waste, Appl. Energy 176 (2016) 116–124, https:// doi.org/10.1016/j.apenergy.2016.05.054.
- [37] M. Okamoto, T. Miyahara, O. Mizuno, T. Noike, Biological hydrogen potential of materials characteristic of the organic fraction of municipal solid wastes, Water Sci. Technol. 41 (2000). http://wst.iwaponline.com/content/41/3/25 (accessed May 23, 2017).
- [38] T. Kobayashi, K.-Q. Xu, Y.-Y. Li, Y. Inamori, Evaluation of hydrogen and methane production from municipal solid wastes with different compositions of fat, protein, cellulosic materials and the other carbohydrates, Int. J. Hydrogen Energy 37 (2012) 15711–15718, https://doi.org/10.1016/ j.ijhydene.2012.05.044.
- [39] J. Pan, R. Zhang, H. El-Mashad, H. Sun, Effect of food to microorganism ratio on biohydrogen production from food waste via anaerobic fermentation, Int. J. 33 (23) (2008) 6968–6975.
- [40] M. Ferchichi, E. Crabbe, G.-H. Gil, W. Hintz, A. Almadidy, Influence of initial pH on hydrogen production from cheese whey, J. Biotechnol. 120 (2005) 402–409, https://doi.org/10.1016/j.jbiotec.2005.05.017.
- [41] J. Foster, Beyond pH homeostasis: the acid tolerance response of Salmonellae. Pummelled with protons, the bacterium launches a strategic acid defense, ASM Am. Soc. Microbiol. News 58 (5) (1992) 266–270.
- [42] S.R. Chaganti, D.H. Kim, J.A. Lalman, Flux balance analysis of mixed anaerobic microbial communities: effects of linoleic acid (LA) and pH on biohydrogen production, Int. J. Hydrogen Energy (2011) 14141–14152, https://doi.org/ 10.1016/j.ijhydene.2011.04.161.
- [43] A. Ghimire, L. Frunzo, F. Pirozzi, E. Trably, R. Escudie, P.N.L. Lens, G. Esposito, A review on dark fermentative biohydrogen production from organic biomass: process parameters and use of by-products, Appl. Energy 144 (2015) 73–95, https://doi.org/10.1016/j.apenergy.2015.01.045.
- [44] X. Wu, W. Yao, J. Zhu, Effect of pH on continuous biohydrogen production from liquid swine manure with glucose supplement using an anaerobic sequencing batch reactor, Int. J. Hydrogen Energy 35 (13) (2010) 6592–6599.
- [45] S.R. Chaganti, B. Pendyala, J.A. Lalman, S.S. Veeravalli, D.D. Heath, Influence of linoleic acid, pH and HRT on anaerobic microbial populations and metabolic shifts in ASBRs during dark hydrogen fermentation of lignocellulosic sugars, Int. J. Hydrogen Energy 38 (2013) 2212–2220, https://doi.org/10.1016/ j.ijhydene.2012.11.137.
- [46] G. Luo, D. Karakashev, L. Xie, Q. Zhou, Long-term effect of inoculum pretreatment on fermentative hydrogen production by repeated batch cultivations: homoacetogenesis and methanogenesis as competitors to hydrogen production, Biotechnol. 108 (8) (2011) 1816–1827.
- [47] B. Calli, J. Zhao, E. Nijssen, K. Vanbroekhoven, Significance of acetogenic H2 consumption in dark fermentation and effectiveness of pH, Water Sci. Technol. 57 (2008) 809–814, https://doi.org/10.2166/wst.2008.089.
- [48] M. Voelklein, A. Jacob, R. O'Shea, J. Murphy, Assessment of increasing loading rate on two-stage digestion of food waste, Bioresour. Technol. 202 (2016) 172–180.
- [49] A. Schievano, A. Tenca, B. Scaglia, Two-stage vs single-stage thermophilic anaerobic digestion: comparison of energy production and biodegradation efficiencies, Sci. Technol. 46 (15) (2012) 8502–8510.
- [50] F. Girotto, M.C. Lavagnolo, A. Pivato, R. Cossu, Acidogenic Fermentation of the Organic Fraction of Municipal Solid Waste and Cheese Whey for Bio-plastic Precursors Recovery €" Effects of Process Conditions during Batch Tests, 2017, https://doi.org/10.1016/j.wasman.2017.09.015.
- [51] W.R.M. Leite, M. Gottardo, P. Pavan, P. Belli Filho, D. Bolzonella, Performance and energy aspects of single and two phase thermophilic anaerobic digestion of waste activated sludge, Renew. Energy 86 (2016) 1324–1331, https:// doi.org/10.1016/j.renene.2015.09.069.
- [52] J. Zhong, D. Stevens, C. Hansen, Optimization of anaerobic hydrogen and methane production from dairy processing waste using a two-stage digestion in induced bed reactors (IBR), Int. J. Hydrog, 40 (45) (2015) 15470–15476.
- [53] C. Li, P. Champagne, B. Anderson, Enhanced biogas production from anaerobic co-digestion of municipal wastewater treatment sludge and fat, oil and grease (FOG) by a modified two-stage, Renew. Energy 83 (2015) 474–482.
- [54] X. Wang, Y. Zhao, A bench scale study of fermentative hydrogen and methane production from food waste in integrated two-stage process, Int. J. Hydrogen Energy 34 (1) (2009) 245–254.
- [55] H. Zhu, A. Stadnyk, M. Béland, P. Seto, Co-production of hydrogen and methane from potato waste using a two-stage anaerobic digestion process, Bioresour. Technol. 99 (2008) 5078–5084, https://doi.org/10.1016/ j.biortech.2007.08.083.
- [56] A. Giordano, C. Cantù, A. Spagni, Monitoring the biochemical hydrogen and methane potential of the two-stage dark-fermentative process, Bioresour. Technol. 102 (2011) 4474–4479, https://doi.org/10.1016/ j.biortech.2010.12.106.
- [57] Z. Wang, S. Shao, C. Zhang, D. Lu, H. Ma, Pretreatment of vinegar residue and anaerobic sludge for enhanced hydrogen and methane production in the twostage anaerobic system, Int. J. 40 (13) (2015) 4494–4501.